



Abundance of coastal fish key functional groups

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1 Key message

This core indicator evaluates the abundance of selected functional groups of coastal fish in the Baltic Sea. As a rule, good status is achieved when the abundance of cyprinids or mesopredators (i.e. mid trophic-level fish) is within an acceptable range for the specific coastal area. The status of functional groups of coastal fish in the Baltic Sea has been evaluated by assessing the status of cyprinids and mesopredators during the period 2016-2020 (Figures 1 and 2).

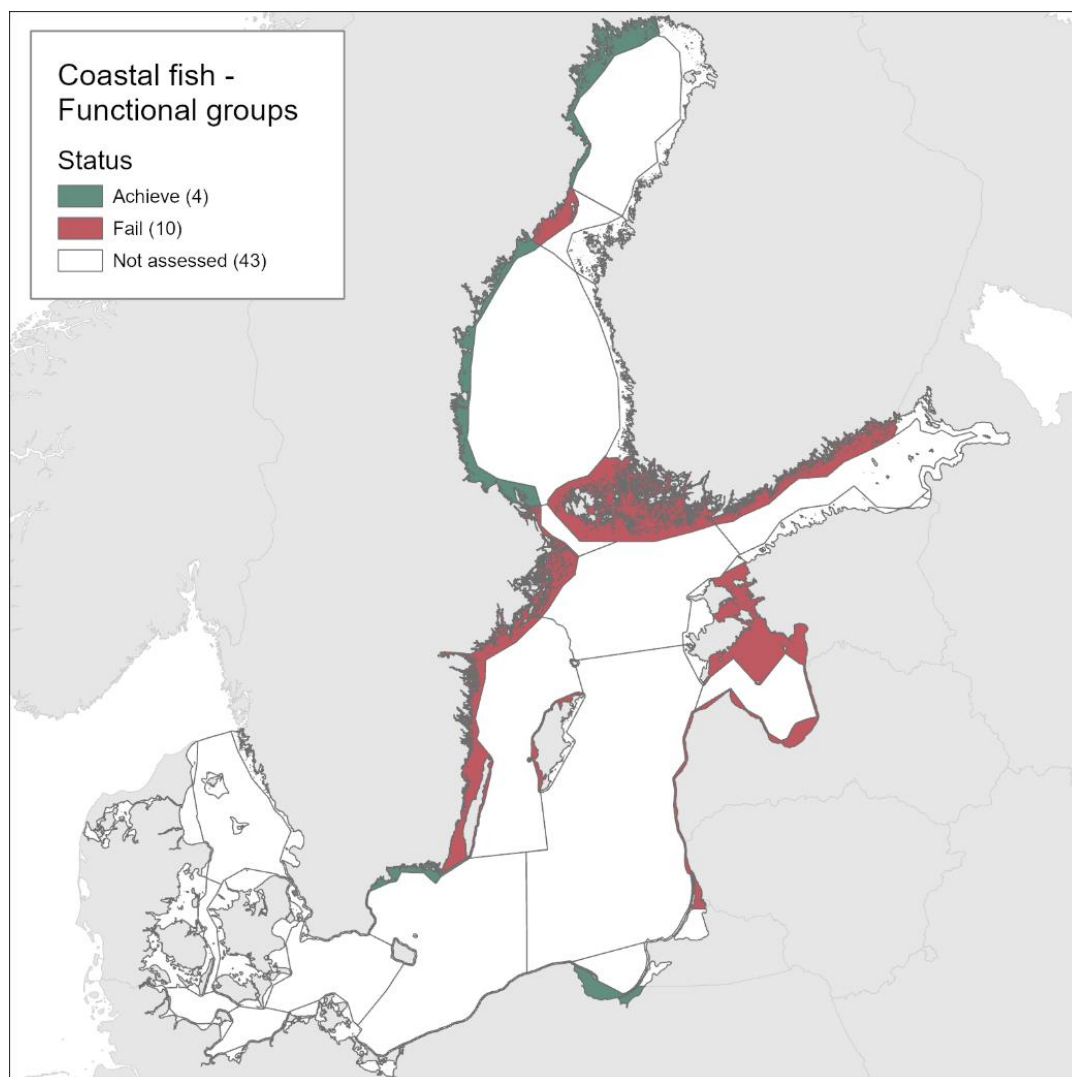


Figure 1. Status evaluation results based on the evaluation of the indicator 'abundance of coastal fish key functional groups' - integrated results of the two functional groups, cyprinids and mesopredators (see Figure 2 for separate). The evaluation is carried out using Scale 3 HELCOM assessment units (defined in the [HELCOM Monitoring and Assessment Strategy Annex 4](#)). See 'data chapter' for interactive maps and data at the [HELCOM Map and Data Service](#).

For cyprinids/mesopredators, good status is achieved in 20 of the 32 monitored locations, but integration of the results of all key species over HELCOM assessment units using the One-Out-All-Out principle, showed that good status achieved in only 4 of the 14 evaluated

assessment units. In the majority of the monitoring locations (24 locations) cyprinids is evaluated, and in 13 of these the threshold is met. For mesopredators the status appears to be better as the threshold is met in 7 of the in total 9 locations evaluated. Note that in one Swedish location (Kvädöfjärden), both cyprinids and mesopredators are evaluated, and neither meets the threshold, and in two Swedish areas included, the time-series is too short to allow for an evaluation of status.

In the locations classified as not good, the abundance of cyprinids and mesopredators was too high in all but two of the 12 locations (i.e. Hiiumaa in Estonia, and Jurkalne in Latvia).

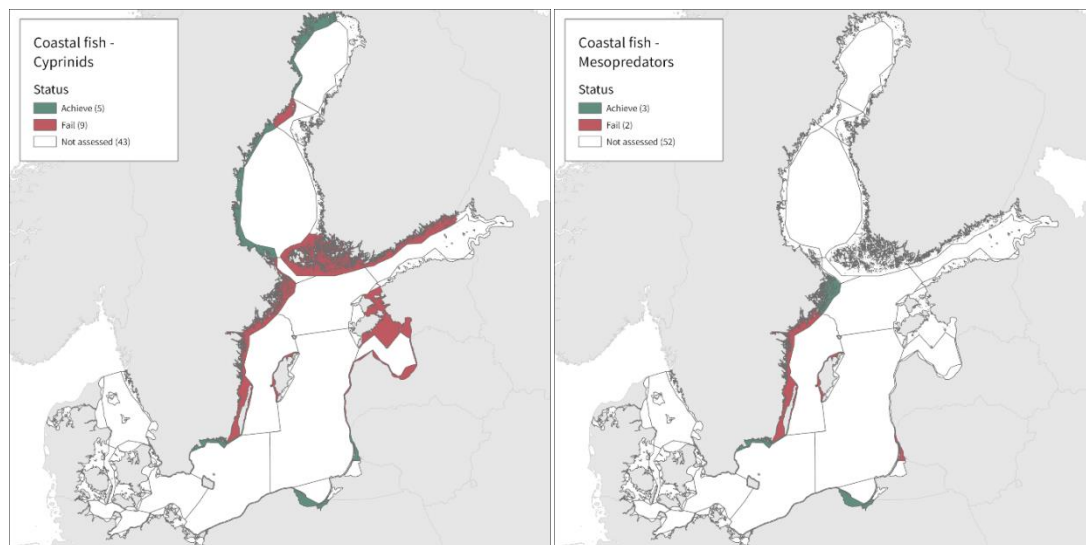


Figure 2. Status evaluation results based on the evaluation of the indicator 'abundance of coastal fish key functional groups' – results shown separately for the two functional groups cyprinids to the left and mesopredators to the right. The evaluation is carried out using Scale 3 HELCOM assessment units (defined in the [HELCOM Monitoring and Assessment Strategy Annex 4](#)). See 'data chapter' for interactive maps and data at the HELCOM Map and Data Service.

Generally, good status is not achieved in more central parts of the Baltic Sea including the Swedish part of the Quark, Åland Sea, Northern Baltic Proper and Western Gotland Basin, in more southern Finnish coastal waters (Archipelago Sea and Gulf of Finland), and in Estonian and Latvian coastal waters. Note that functional groups are not evaluated in the Finnish coastal areas of the Bothnian Bay and Bothnian Sea due to lack of data.

The level of confidence in the evaluation differs between areas and regions due to differences in monitoring methodology as well due to lower temporal and spatial coverage of monitoring in some countries, the latter generally relating to resource availability. The methodological confidence is high in all areas, and the confidence in the accuracy of the evaluation is high in the majority of the assessment units. The confidence in the temporal coverage is high in all areas except for Latvian and Lithuanian coastal areas, and the confidence in spatial representability is moderate to high in all assessment units evaluated besides those in Estonia and Latvia. The overall integrated confidence evaluation considering all four categories is high in five assessment units and intermediate in the remaining nine units, with no clear spatial pattern.

The indicator is operational in the coastal waters of most countries bordering the Baltic Sea. For the time being, it is not applicable in some areas where coastal fish monitoring data are scarce, or where the group meso-predators overlaps heavily with the species reported under the indicator "Abundance of coastal fish key species" and further studies as well as time series are needed to yield a reliable evaluation. In the future, in line with increasing knowledge, the indicator might undergo further development.

1.1 Citation

The data and resulting data products (e.g. tables, figures and maps) available on the indicator web page can be used freely given that it is used appropriately and the source is cited. The indicator should be cited as follows:

HELCOM (2023) Abundance of coastal fish key functional groups. HELCOM core indicator report. Online. [Date Viewed], [Web link].

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2 Relevance of the indicator

The state of coastal fish communities reflects the ecological state of coastal ecosystems, and in some areas where cyprinids and mesopredators are targeted, the effects of mainly small-scale coastal commercial fisheries. Changes in the long-term development of the abundance of coastal fish functional groups reflects the effects of increased water temperature and eutrophication in coastal areas, and/or changes in the level of human exploitation (mainly habitat degradation), natural predation pressure, and in some areas fishing.

2.1 Ecological relevance

Coastal fish are recognized as being important components of coastal food webs and ecosystem functioning and high abundances of cyprinids and mesopredatory fish are generally indicative of poorer environmental conditions in the coastal ecosystem (Eriksson *et al.* 2009; Baden *et al.* 2012; Bergström *et al.* 2016b, 2019; Östman *et al.* 2016). High abundances of cyprinids and mesopredators might reflect lack of top-down regulation, elevated eutrophication and increased water temperatures. In Sweden and Finland, a fishery targeting cyprinids has developed during recent years (Lappalainen *et al.* 2019; Dahlin *et al.* 2021), and resulting effects on targeted populations might hence be seen in the future.

Moreover, since many coastal fish species are rather local in their appearance (Saulamo & Neuman 2005; Laikre *et al.* 2005; Olsson *et al.* 2011; Östman *et al.* 2017a), the temporal development of coastal fish communities might reflect the general environmental state in the monitoring locations (Bergström *et al.* 2016b, 2019; Östman *et al.* 2017b).

2.2 Policy relevance

The core indicator on abundance of coastal fish functional groups addresses the Baltic Sea Action Plan's ([BSAP 2021](#)). Biodiversity and nature conservation segment's ecological objectives 'Natural distribution and occurrence of plants and animals' and 'Thriving and balanced communities of plants and animals'.

The core indicator is relevant to the following specific BSAP actions:

- 'to develop long-term plans for, protecting, monitoring and sustainably managing coastal fish species, including the most threatened and/or declining, including anadromous ones (according to the HELCOM Red list of threatened and declining species of lampreys and fishes of the Baltic Sea, BSEP No. 109), by 2012' and
- 'develop a suite of indicators with region-specific reference values and targets for coastal fish as well as tools for evaluation and sustainable management of coastal fish by 2012'.

The core indicator also addresses the following qualitative descriptors of the MSFD for determining good environmental status:

Descriptor 4: 'All elements of the marine food webs, to the extent that they are known, occur at normal abundance and diversity and levels capable of ensuring the long-term abundance of the species and the retention of their full reproductive capacity'.

and the following criteria of the Commission Decision:

- Criterion D4C2 (Trophic guilds, balance of total guild abundance).

In some Contracting Parties the indicator also has potential relevance for implementation of the EU Habitats Directive.

A summary is provided in Table 1.

Table 1. Policy relevance

	Baltic Sea Action Plan (BSAP)	Marine Strategy Framework Directive (MSFD)
Fundamental link	Segment: Biodiversity Goal: "Baltic Sea ecosystem is healthy and resilient" <ul style="list-style-type: none"> • Ecological objective: "Functional, healthy and resilient food webs". • Management objective: "Reduce or prevent human pressures that lead to imbalance in the food web". 	Descriptor 4 Ecosystems, including food webs - Trophic guilds of an ecosystem <ul style="list-style-type: none"> • Criteria 2 The balance of total abundance between the trophic guilds is not adversely affected due to anthropogenic pressures. • Feature – Coastal ecosystems. • Element of the feature assessed – Coastal fish species.
Complementary link	Segment: Biodiversity Goal: "Baltic Sea ecosystem is healthy and resilient" <ul style="list-style-type: none"> • Ecological objective: "Viable populations of all native species". • Management objective: "Human induced mortality, including hunting, fishing, and incidental bycatch, does not threaten the viability of marine life". 	
Other relevant legislation:	In some Contracting Parties of HELCOM - potentially also EU Habitats Directive. UN Sustainable Development Goal 14 (Conserve and sustainably use the oceans, seas and marine resources for sustainable development) is most clearly relevant, though SDG 12 (Ensure sustainable consumption and production patterns) and 13 (Take urgent action to combat climate change and its impacts) also have relevance.	

2.3 Relevance for other assessments

The status of biodiversity is assessed using several core indicators. Each indicator focuses on one important aspect of the complex issue. In addition to providing an indicator-based evaluation of the abundance of selected functional groups of coastal fish, this indicator also contributes to the overall biodiversity assessment along with the other biodiversity core indicators.

3 Threshold values

Good status is achieved when the abundance of cyprinids/mesopredators is within an acceptable range. The quantitative threshold values for coastal fish are based on location-specific reference conditions where time series covering more than 15 years are available (ten or more years reference period + five or more years assessment period). In areas where shorter time series (i.e. less than 15 years) are available, a trend-based approach is used.

A reference period needs to be defined for determining the threshold value. The period used to define the reference needs to cover at least ten years in order to extend over more than twice the generation time of the typical species represented in the indicator and thus cater for natural variation in the indicator value, due for example to strong and weak year classes. For the period used to determine the reference to be relevant, it must also be carefully selected to reflect time periods with stable environmental conditions, as stated within the MSFD (European Commission 2008). Substantial turnovers in ecosystem structure in the Baltic Sea were apparent in the late 1980s, leading to shifts in the baseline state (Möllmann *et al.* 2009), and for coastal fish communities, substantial shifts in community structure have been demonstrated in the late 1980s and early/mid 1990s (Olsson *et al.* 2012; Bergström *et al.* 2016a). In some areas, there have also been minor shifts in fish community structure later. To account for this, the ASCETS method (Östman *et al.* 2020) is applied on time-series with more than 15 years of data. This method offers a refined approach to infer structural changes in indicator values over time and establish threshold values for the state during a reference period based on the observed variation in indicator values.

Estimates of the relative abundance and/or biomass are used to determine whether coastal fish key functional groups in the Baltic Sea achieve good status or not. These estimates are derived from fishery independent monitoring. Since there are strong environmental gradients in the Baltic Sea and coastal fish communities, stocks are typically local in their appearance and respond mainly to area-specific environmental conditions. The evaluations for coastal fish key functional groups are thus carried out on a relatively local scale.

The assessment period applied when using the ASCETS method should cover at least five years to cater for natural variability. Good status is evaluated based on the deviation of the median value of the indicator during the assessment period in relation to the threshold value (Figure 3).

Cyprinids/mesopredators

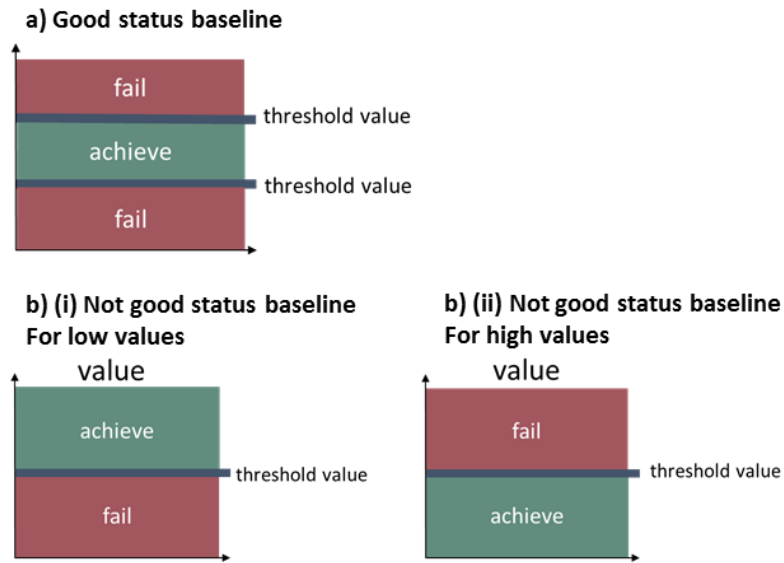


Figure 3. Determination of acceptable range from baseline.

When using the trend-based approach, environmental status is evaluated based on the direction of the trend towards good status, over the time period 2014-2020 (Figure 4).

Cyprinids/mesopredators

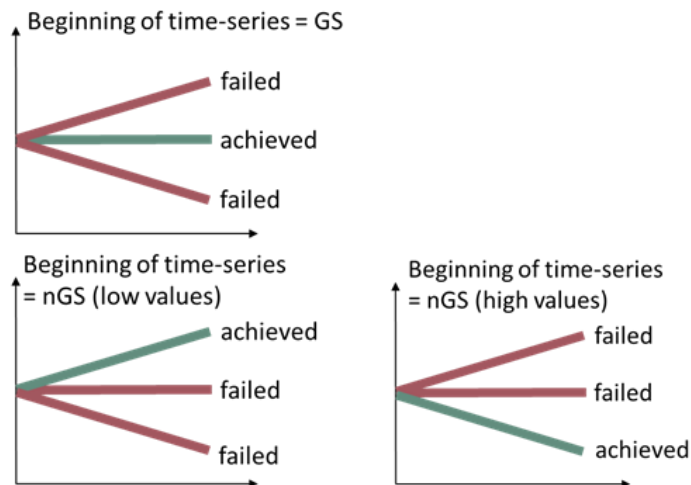


Figure 4. Application of the trend-based approach for evaluating environmental. The status is defined based on the direction of the trend of the indicator compared to the desired direction of the indicator over time. GS = good status, nGS = not good status.

The functional groups used in this indicator are members of the cyprinid family. In areas where cyprinids do not exist naturally, mesopredatory fish species are used e.g. any mid-trophic level species that are not piscivorous. The composition of cyprinid and mesopredator species differ along the coast. The most abundant species in the Cyprinid family (*Cyprinidae*) in the less saline eastern and northern parts of the Baltic Sea are for example roach (*Rutilus rutilus*) and bream (*Abramis sp.*), whereas mesopredatory fish are representative of the more exposed coastal parts of the central Baltic Sea and in its more saline western region.

Table 2. Species included in the two functional groups cyprinids and mesopredators in the different countries for which the indicator is currently applicable. Presence is indicated according to the following; X: Occurs in monitoring in representative numbers, X*: Occurs in monitoring in representative numbers, but no identification of the different species is possible, x: Occurs in monitoring but in low and non-representative numbers, blank: Not applicable in the country. Countries: FI: Finland, EE: Estonia, LV: Latvia, LT: Lithuania, PL: Poland, SE: Sweden.

Cyprinids	FI	EE	LV	LT	PL	SE
Roach (<i>R. Rutilus</i>)	X	X	X	X	X	X
Rudd (<i>S. Erythrophthalmus</i>)	X	X	X	x	x	X
Bleak (<i>A. Albumus</i>)	X	X	X	X	x	X
Common bream (<i>A. Brama</i>)	X	x	X	X	X	X
White bream (<i>A. Bjoerkna</i>)	X	X	X	X	X	X
Zope (<i>A. Ballerus</i>)					x	
Wimba bream (<i>V. vimba</i>)	X	X	X	X	x	x
Ide (<i>L. Idus</i>)	X	X	X	x	x	x
Dace (<i>L. Leusicus</i>)		x	X	x		x
Crucian carp (<i>C. Carassius</i>)	x	X	X	x	X	x
Gibel carp (<i>C. Gibelio</i>)		X				
Tench (<i>T. Tinca</i>)	x	x		x	x	x
Minnow (<i>P. Phoxinus</i>)		x				x
Gudgeon (<i>G. Gobio</i>)		X				
Chub (<i>S. cephalus</i>)		x			x	
Sichel (<i>P. cultratus</i>)	x	x			x	
Mesopredators	FI	EE	LV	LT	PL	SE
All cyprinid fish (see above)	X	X	X	X	X	X
Flounder (<i>P. Flesus</i>)	X*	X*	X*	X*	X	X*
Baltic flounder (<i>P. Solemdali</i>)	X*	X*	X*	X*		X*
Ruffe (<i>G. Cernuus</i>)	X	X	X	X	x	X
Eel (<i>A. Anguilla</i>)		x			x	X
Herring (<i>C. Harengus</i>)	X	X	X	X	x	X
Sprat (<i>S. Sprattus</i>)	X	x	X	X	x	x
Smelt (<i>O. Eperlanus</i>)	X	X	X	X	x	x
Plaice (<i>P. Platessa</i>)					x	x
Common dab (<i>L. Limanda</i>)						
Common sole (<i>S. Solea</i>)						x
Whitefish (<i>C. Maraena</i>)	X	X	X	X	x	X
Eelpout (<i>Z. Viviparus</i>)	X	X	X	x	x	X
Vendace (<i>C. Albula</i>)	x	x				X
Labrids (<i>L. Berggylta</i> , <i>L. Mixtus</i> , <i>C. Exoletus</i> , <i>S. Melops</i> , <i>C. Rupestris</i>)						X
Sculpins (<i>C. Poecilopus</i> , <i>T. Quadricornis</i> , <i>T. Bubalis</i> , <i>A. Cataphractus</i> , <i>M. Scorpius</i>)	X	X	X	X	x	X
Gobies (<i>G. Niger</i> , <i>N. Melanostomus</i>)	X	X	X	X	X	X
Sticklebacks (<i>G. Aculeatus</i> , <i>P. Pungitius</i>)	X	x	X	x	x	X
Rocklings (<i>C. Mustela</i> , <i>E. Cimbrius</i>)						x
Pipefishes (<i>E. Aequoreus</i> , <i>S. Acus</i> , <i>S. Rostellatus</i> , <i>S. Tyhple</i>)	X	x	x	x	x	x
Garfish (<i>B. Belone</i>)		x		x	x	
Lumpfish (<i>C. Lumpus</i>)		x		x		x
Lesser sand-eel (<i>A. Marinus</i>)					x	
Small sandeel (<i>A. tobianus</i>)	x	x	X	x	x	x
Great sandeel (<i>H. lanceolatus</i>)	x	x	X	x	x	x

3.1 Setting the threshold value(s)

To determine the status of the indicator, the ASCETS method first derives a bootstrapped distribution of median values from a time series of observed indicator values during a reference period. Specific threshold values for changes in indicator state is set, and for key species, these are based on the 5th and 98th percentile values of the bootstrapped distribution. In this way, the derived boundaries of this interval can function as threshold values for a change in state per assessment unit of each species. Second, the bootstrapped median indicator value during the assessment period is evaluated in relation to the threshold values derived from the reference period depending on how much of the bootstrapped median distribution from the assessment period that falls below, within, or above the 5th and 98th percentiles.

If the requirements for defining quantitative baseline conditions are not met (e.g. short time-series), then a trend-based evaluation should be used. All available data starting from year 2014 is included in trend analyses. In the trend-based approach, good status is defined based on the direction of the trend at $p < 0.1$ of the indicator compared to the desired direction of the indicator over time.

4 Results and discussion

The results of the indicator evaluation that underlie the key message map and information are provided below.

4.1 Status evaluation

The current evaluation of coastal fish environmental status covers the period 2016-2020. The evaluation is based on time-series data of varying length depending on the temporal coverage of data collection in each monitoring location. Time series thus start between the years 1998 and 2015 (Table 2) and depending on the time-series coverage, either the 'ASCETS approach' or a 'trend-based evaluation' is used. Evaluations were carried out for 14 of the in total 42 scale 3 assessment units and time series data up to and including the year 2020 were available for all 14 of these units.

The environmental status of cyprinids and mesopredator abundance is generally not good. Good status is achieved in 63 % of the evaluated monitoring locations (20 out of in total 32 locations), but only 4 out of 14 assessment units achieve good status (see Table 3). In the locations classified as not good, the abundance of cyprinids and mesopredators was too high in all but two (Hiiumaa, Estonia, and Jurkalne, Latvia) of the 12 locations.

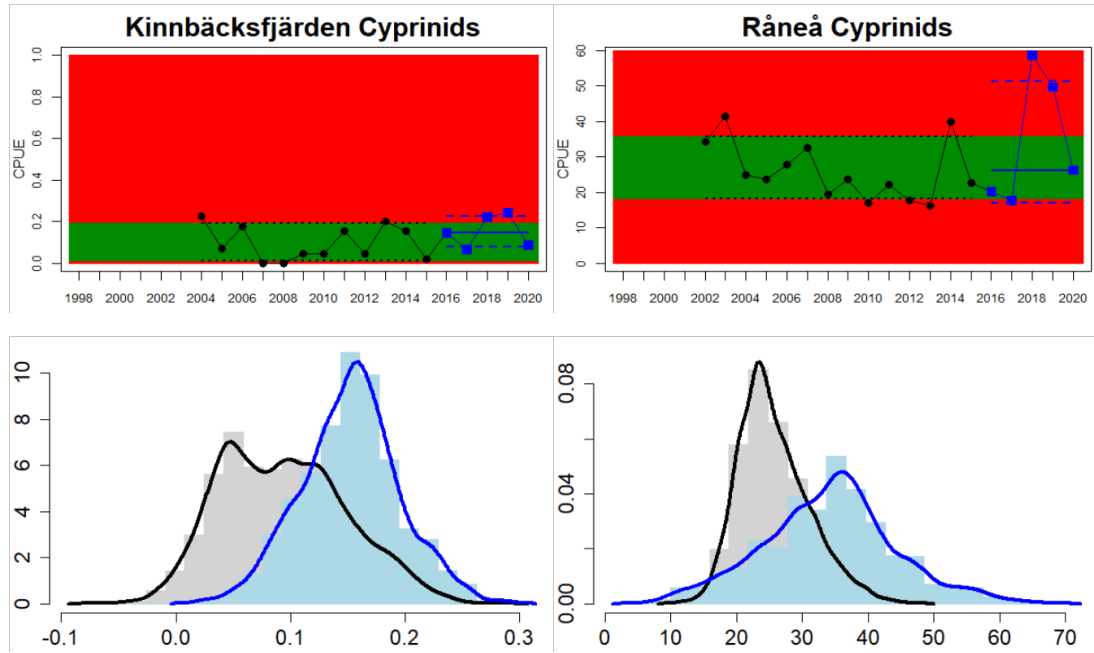
In the majority of the monitoring locations (24 locations) cyprinids is evaluated, and in 13 of these the threshold is met. For mesopredators the status appears to be better as the threshold is met in 7 of the in total 9 locations evaluated. Note that in one Swedish location (Kvädöfjärden), both cyprinids and mesopredators are evaluated, and neither meets the threshold, and in two Swedish areas included, the time-series is too short to allow for an evaluation of status.

There are some geographical patterns in the status of the cyprinids/mesopredators, and good status is generally not achieved in more central parts of the Baltic Sea including the Swedish part of the Quark, Åland Sea, Northern Baltic Proper and Western Gotland Basin, in more southern Finnish coastal waters (Archipelago Sea and Gulf of Finland), and in Estonian and Latvian coastal waters.

Evaluations of the indicator were only carried out for cyprinids/mesopredators in the central and northern parts of the Baltic Sea since monitoring to support the indicator is currently lacking in Germany and Denmark, and in the Northern parts of Finland (Bothnian Bay and Bothnian Sea). Coastal fish monitoring is not available in Russia.

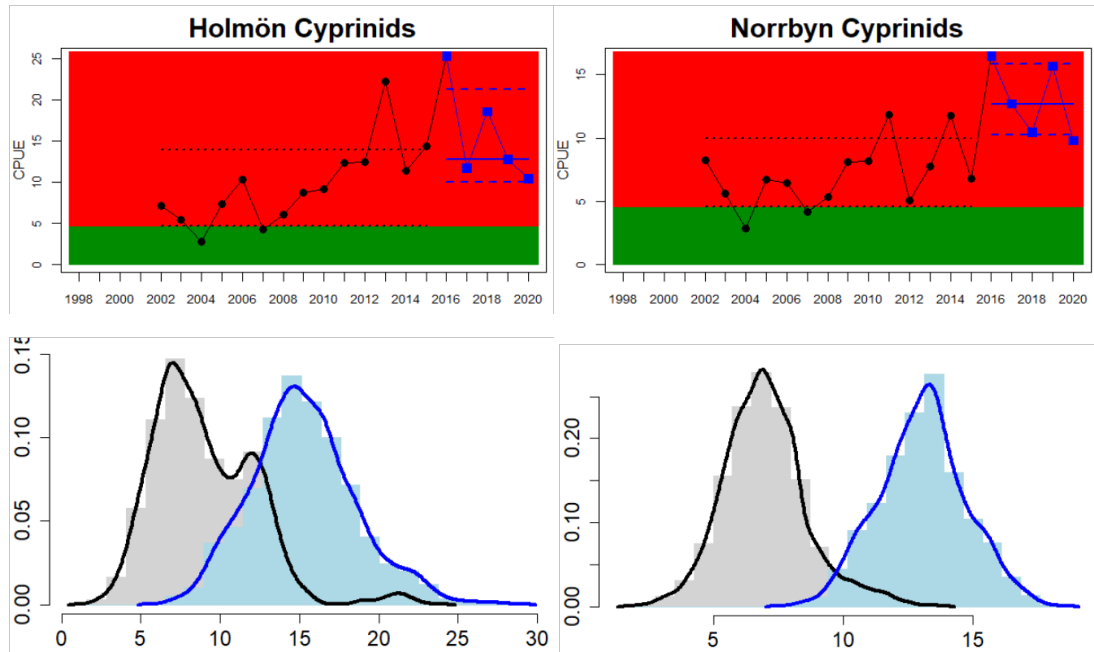
Bothnian Bay

Sweden



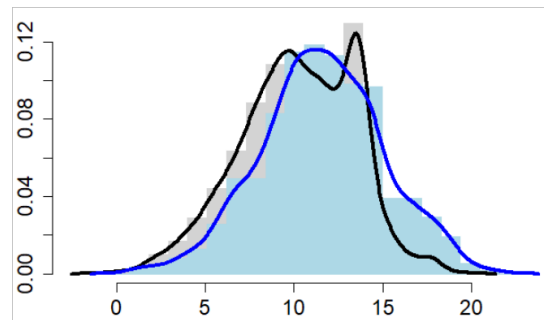
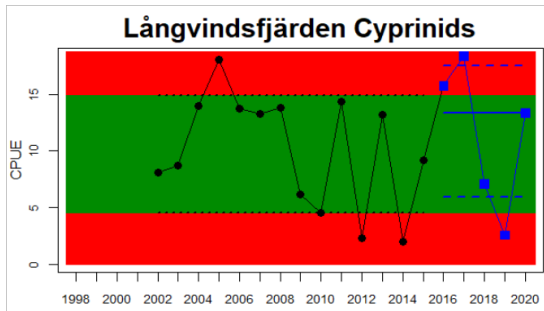
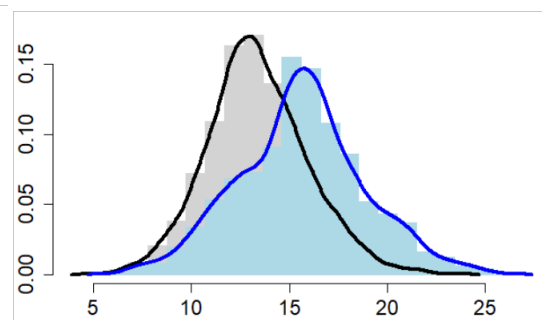
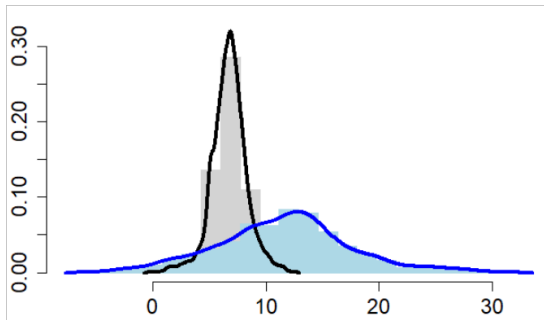
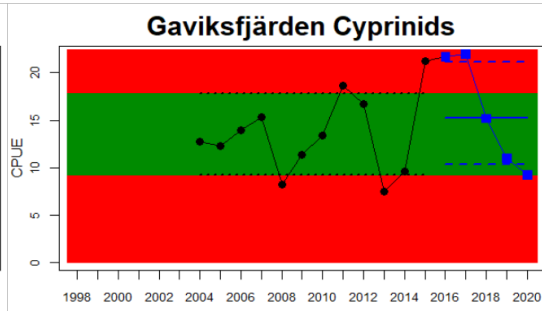
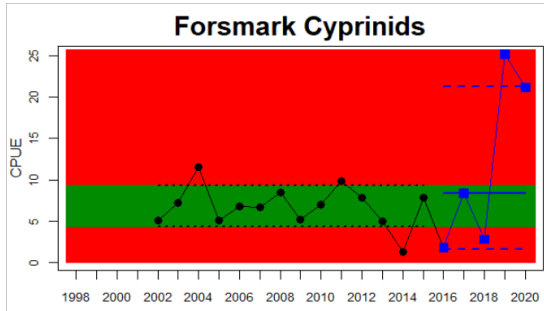
The Quark

Sweden



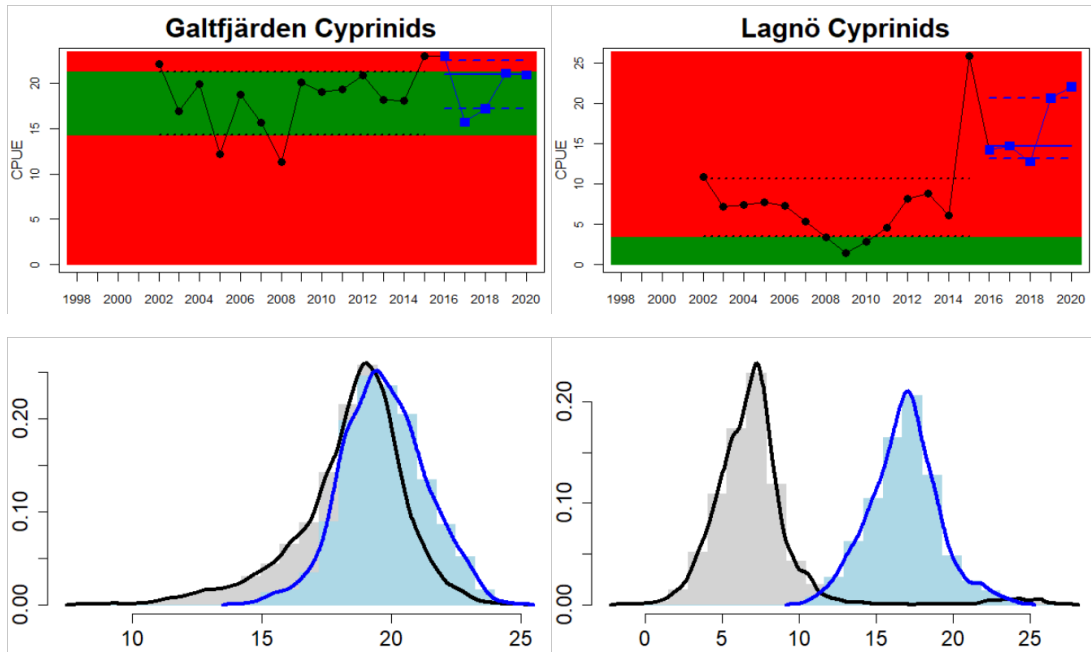
Bothnian Sea

Sweden



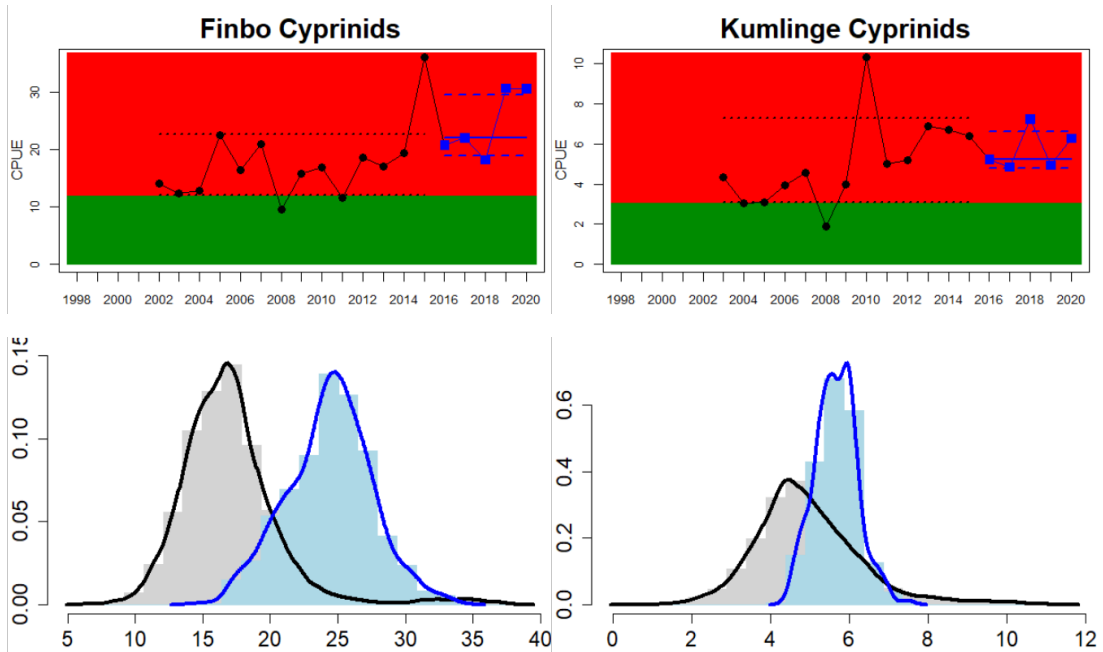
Åland Sea

Sweden



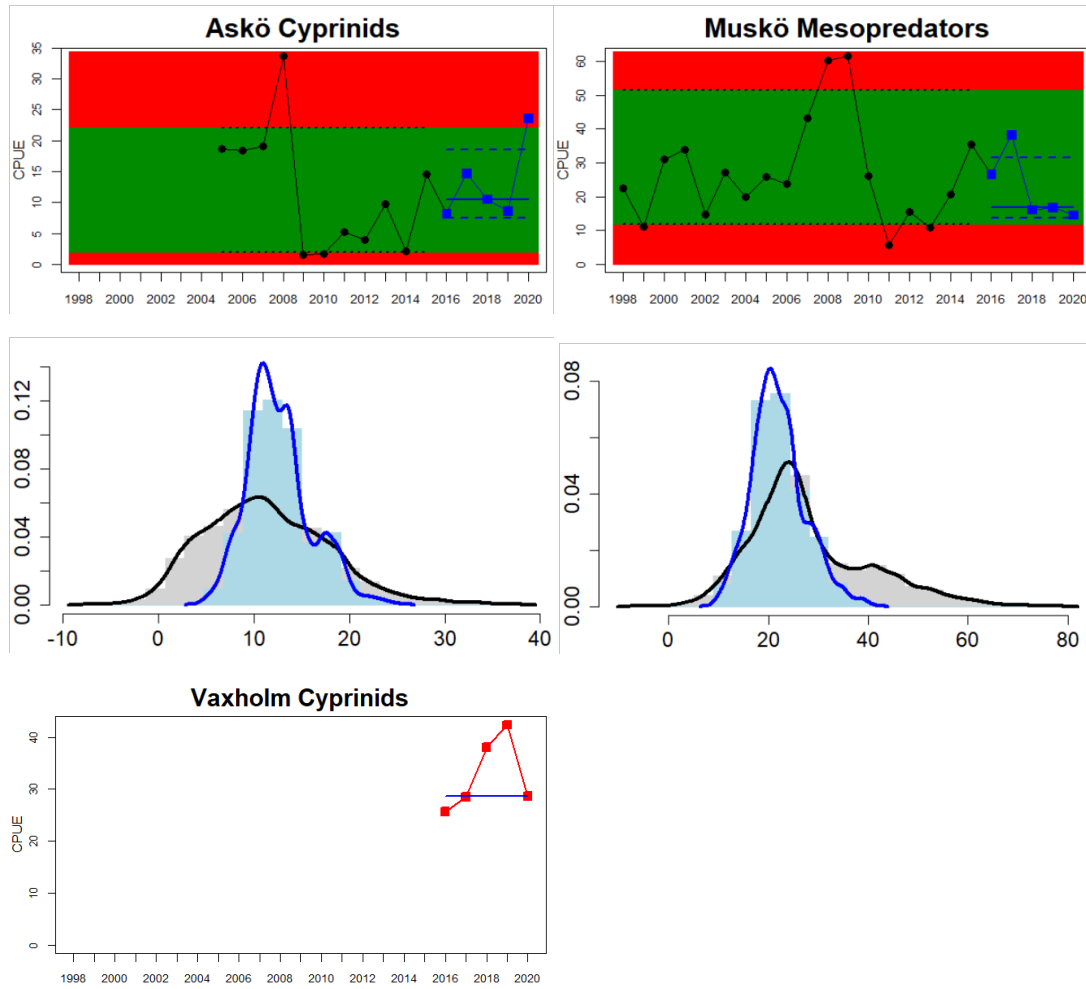
Archipelago Sea

Finland



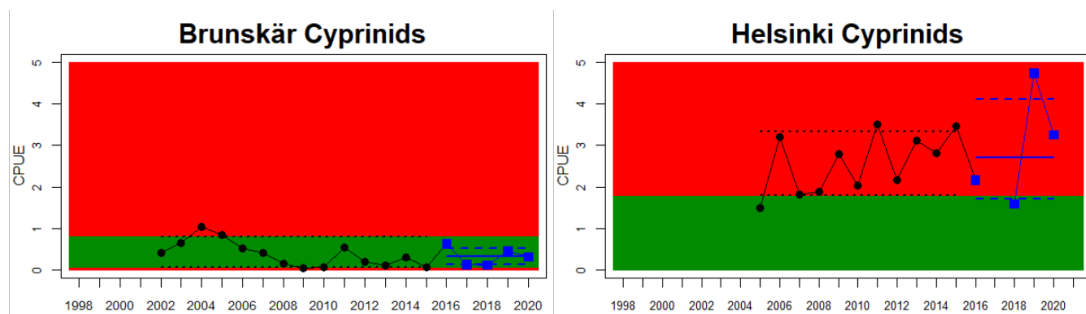
Northern Baltic Sea

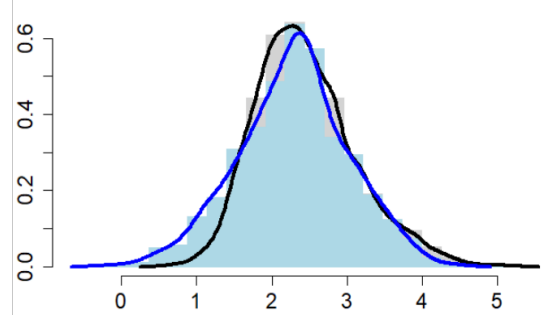
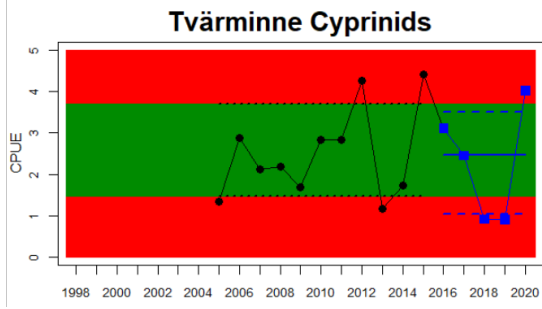
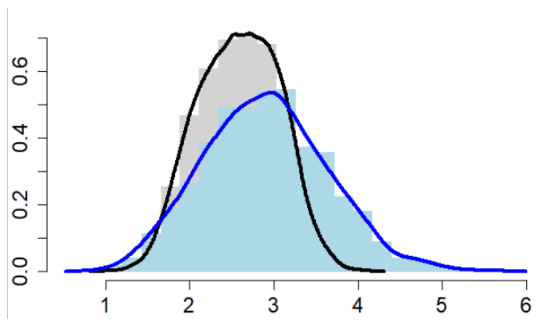
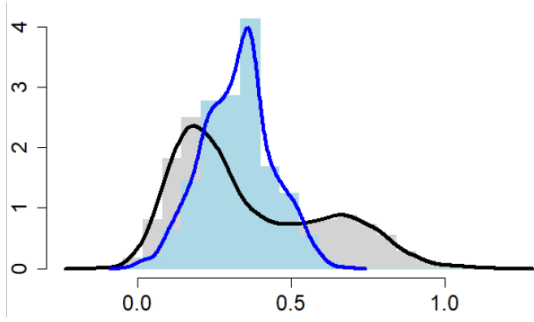
Sweden



Gulf of Finland

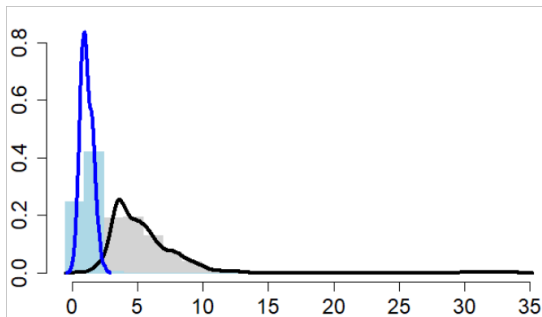
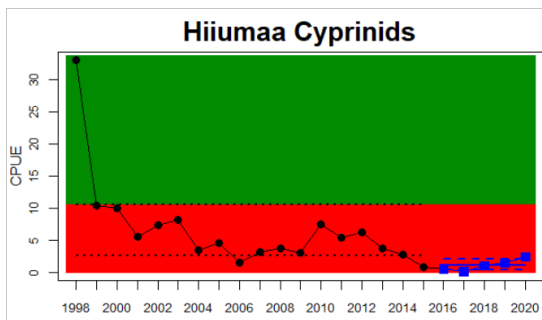
Finland



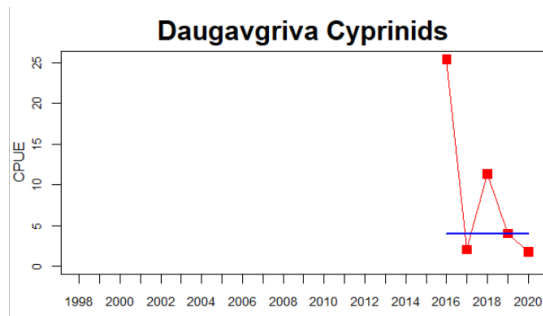


Gulf of Riga

Estonia

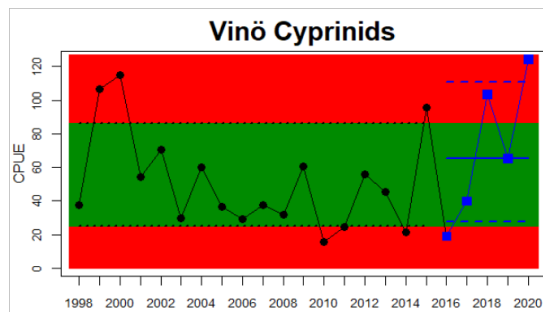
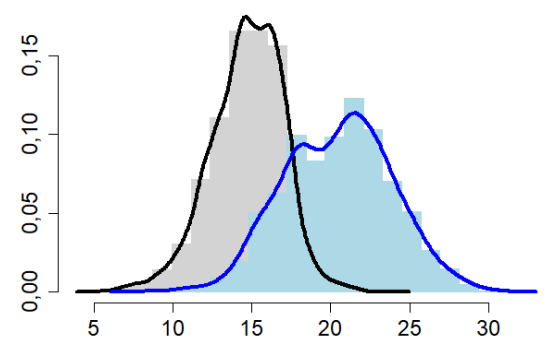
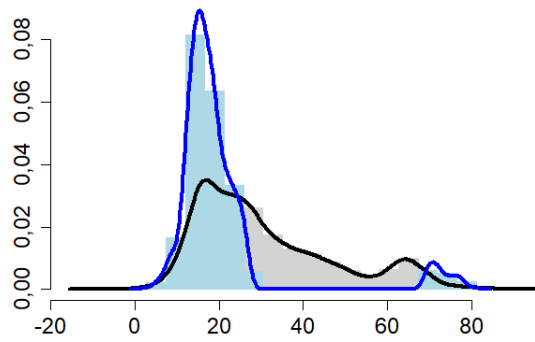
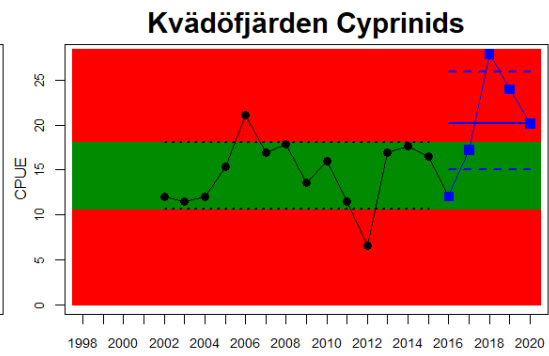
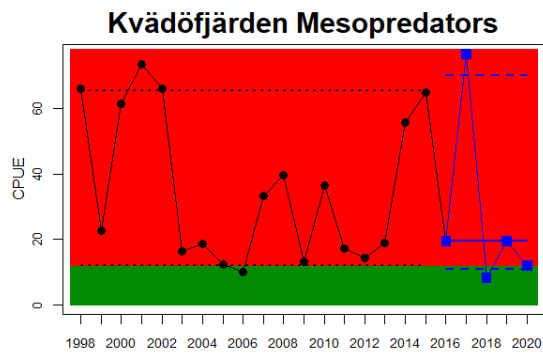


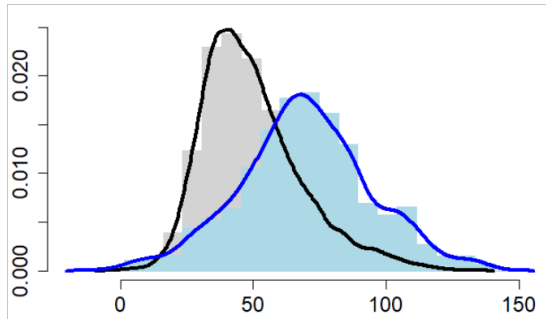
Latvia



Western Gotland Basin

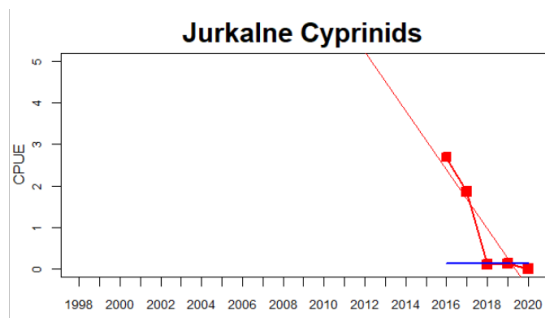
Sweden



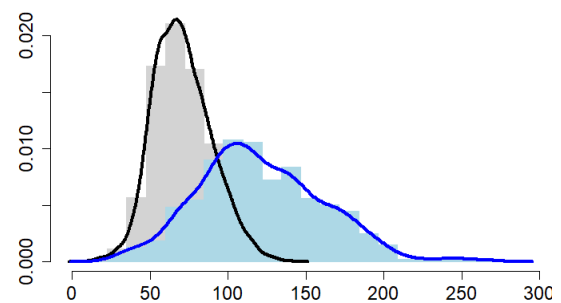
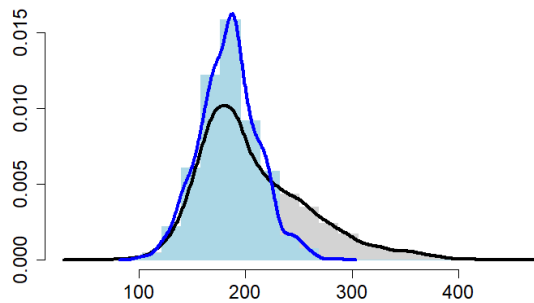
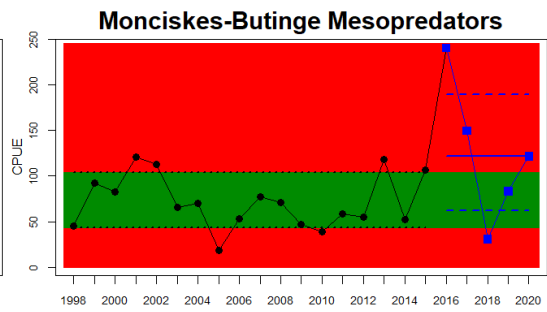
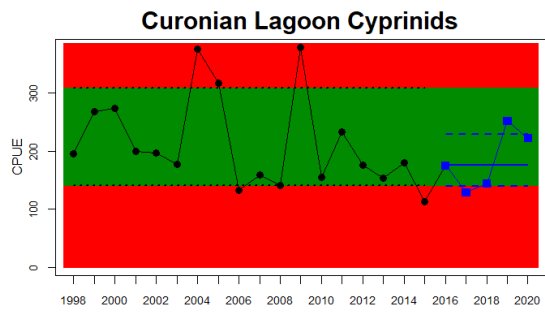


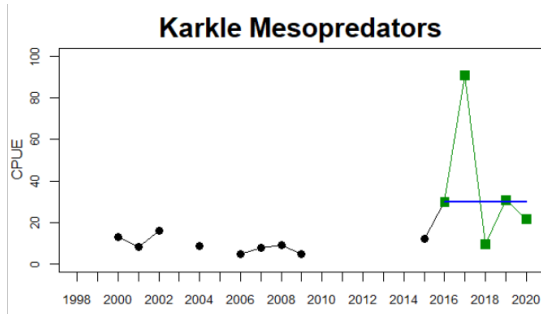
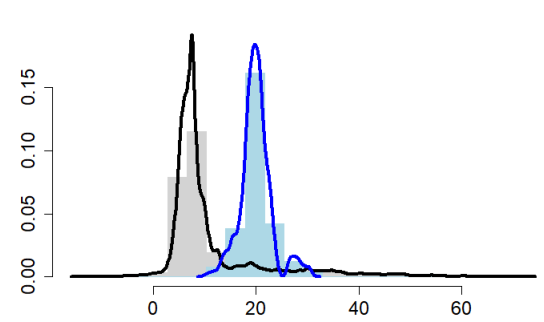
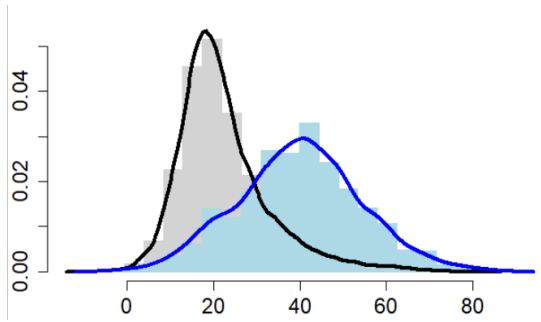
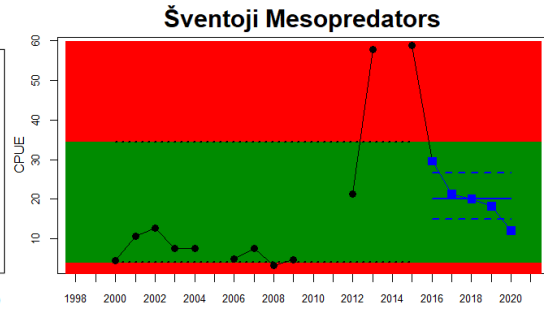
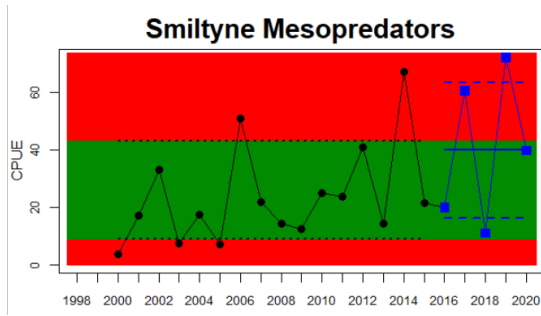
Eastern Gotland Basin

Latvia

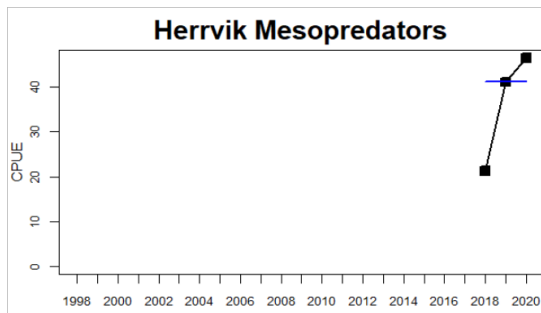


Lithuania



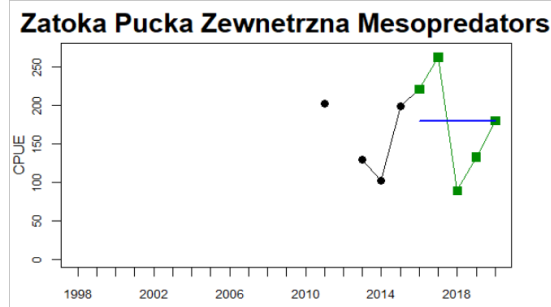
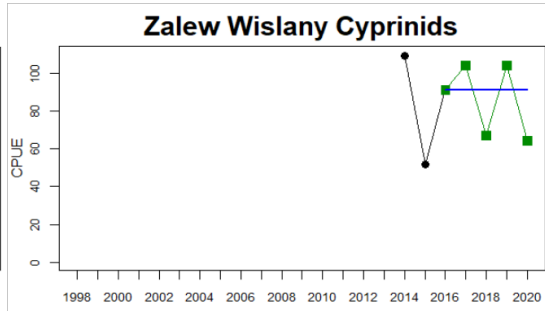
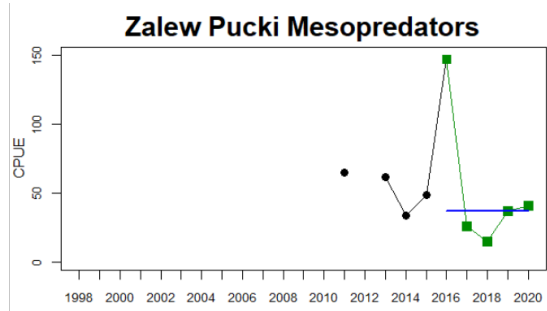


Sweden



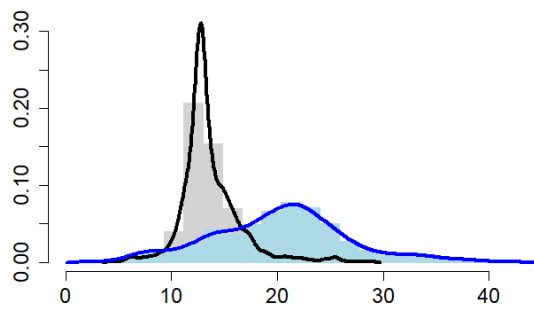
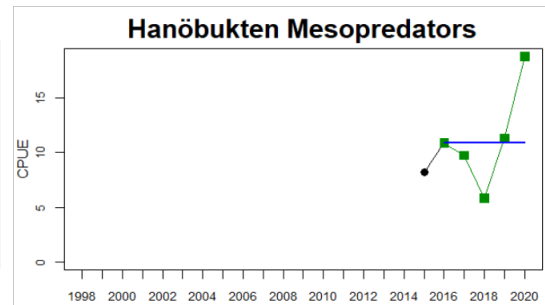
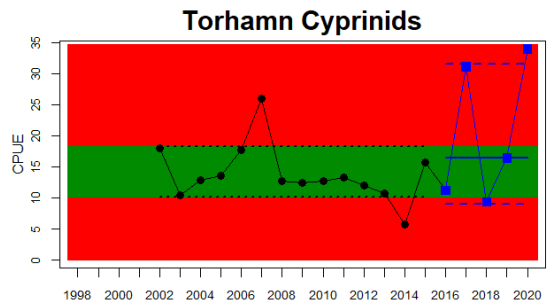
Gdansk Basin

Poland



Bornholm Basin

Sweden



Arkona Basin

Sweden

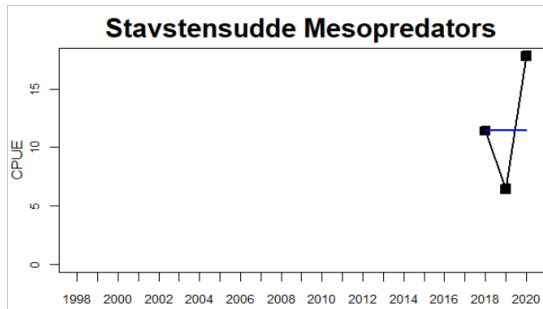


Figure 5. Cyprinid/mesopredator evaluation outcome. All evaluations are displayed per sub-basin and country for each monitoring location. In locations where the ASCETS approach is applied, threshold values are displayed by black dotted lines between fields in green (good status) and red (not good status), with the colour of the fields determined by the status during the reference period. The evaluation of good status/not good status is performed for the assessment period compared to the reference period by comparing the location of the median during the assessment period (full blue line) with the location of the respective threshold line. The 95th percentile intervals associated with the median displayed in hatched blue lines. Below each ASCETS graph, a small graph shows the smoothed bootstrapped medians of the indicator values from the reference period (bars in grey with a black line) and the assessment period (bars in blue with a blue line). For assessment units where the available data only allowed for a trend-based evaluation, green squares denote a good status evaluation outcome during the assessment period whereas red squares denote a not good status evaluation outcome. The hatched trend-line indicates a significant positive (green) and negative (red) trend at $p < 0.1$ during 2014-2020 for the times-series in each location.

4.2 Trends

There is a tendency for a slight decrease in the status of coastal fish in the Baltic Sea when considering cyprinids and mesopredators between this evaluation and HOLAS 2, conducted in 2018 including data until 2016 (Table 3). In three of the assessment units also considered in HOLAS 2, the status has decreased, and in the remaining ten assessment units there is no change over time in status. However, the decreased overall status partly reflects the inclusion of additional areas and functional groups (mesopredators) in some assessment units and areas (see comments in Table 3). The use of a stricter integrating approach across monitoring locations (majority rule in HOLAS 2 vs One-Out-All-Out principle in the current evaluation), might also contribute to the pattern observed.

Table 4. Overview of trends between current and previous evaluation in year 2018 (HOLAS 2, including data until 2016). For each HELCOM assessment unit, it is noted whether the integrated status using the BEAT tool achieves or fails to achieve the threshold value. The current integrated status is compared to the previous status with regards to any distinct increasing or decreasing trend. In case of changed integrated status, the outcome is briefly described focusing on the relevant changes compared to the previous evaluation.

HELCOM Assessment unit name	Threshold value: achieved/failed	Distinct trend between current and previous evaluation	Description of outcomes
Archipelago Sea Coastal waters	failed	no change	
Bornholm Basin Swedish Coastal waters	achieved	no change	
Bothnian Bay Finnish Coastal waters	NA	NA	Included in HOLAS 2, but not in HOLAS III
Bothnian Bay Swedish Coastal waters	achieved	no change	
Bothnian Sea Finnish Coastal waters	NA	NA	Included in HOLAS 2, but not in HOLAS III
Bothnian Sea Swedish Coastal waters	achieved	no change	
Eastern Gotland Basin Latvian Coastal waters	failed	no change	
Eastern Gotland Basin Lithuanian Coastal waters	failed	decrease	Inclusion of 3 new monitoring locations, all with GS, but status is decreased due to nGS in Mon/But
Gulf of Finland Finnish Coastal waters	failed	no change	
Gdansk Basin Polish Coastal waters	achieved	NA	Not included in HOLAS 2
Gulf of Riga Estonian Coastal waters	failed	no change	
Gulf of Riga Latvian Coastal waters	failed	no change	
Northern Baltic Proper Swedish Coastal waters	failed	decrease	Inclusion of two new monitoring locations, status has decreased due to inclusion of Vaxholm
The Quark Finnish Coastal waters	NA	NA	Included in HOLAS 2, but not in HOLAS III
The Quark Swedish Coastal waters	failed	no change	
Western Gotland Basin Swedish Coastal waters	failed	decrease	Due to inclusion of mesopredators in Kvädöfjärden, status has decreased
Åland Sea Swedish Coastal waters	failed	no change	

4.3 Discussion text

The overall environmental status of coastal fish in the Baltic Sea is poor. When summarising the results across cyprinids and mesopredators good status is only achieved in 4 out of the 14 assessment units analysed. 32 monitoring locations are considered in total, and among these, good status is achieved in 20 locations only. In the locations classified as not good, the abundance of cyprinids and mesopredators was too high in all but two (Hiiumaa, Estonia, and Jurkalne, Latvia) of the 12 locations.

There are some geographical patterns in the status of the cyprinids/mesopredators. Good status is only achieved along the Swedish coasts of the Bothnian Bay and Bothnian Sea, as well as along the southernmost Swedish coast (Bornholm Basin) and Polish coastal areas (Gdansk Basin).

5 Confidence

In general, the confidence varies between assessment units, countries and monitoring programmes since, for example, the number of years for which coastal fish monitoring has been carried out varies between locations, as does the spatial coverage of monitoring within assessment units. Generally, the confidence of the evaluation is higher in locations where monitoring started before 1999 and where data is available for all years during the assessment period (2016-2020) and where there is good spatial coverage of monitoring.

The confidence scoring followed the principles as outlined in the HELCOM integrated biodiversity assessment. Confidence was scored using four criteria with three different levels (1 = high, 0.5 = intermediate, and 0 = low). The criteria used were:

Confidence in the accuracy of the estimate (ConfA). In the ASCETS approach, confidence in the evaluation is determined by the C(S) value. C(S) varies between 0 and 1, with values <0.1 representing high confidence of changed status and values >0.9 high confidence of unchanged status (Level 1). Values of 0.1-0.3 represent medium confidence in changed status and 0.7-0.9 medium confidence in unchanged status (Level 0.5). Values of 0.3-0.5 represent low confidence of changed status and 0.5-0.7 low confidence in unchanged status (Level 0). In the trend-based approach, confidence in the evaluation is determined by the p-value of the linear regression, with p-values <0.05 representing high confidence in a trend, p<0.1 medium confidence in a trend, p 0.10-0.20 low confidence in no trend, p 0.21-0.49 medium confidence in no trend, and p 0.5-1.0 high confidence in no trend.

Confidence in the temporal coverage of evaluation (ConfT). Level 1 = data for all years during 2016-2020, 0.5 = data missing for one or two years during 2016-2020, and 0 = data missing for three or more years during 2016-2020.

Confidence in spatial representability of the evaluation (ConfS). Level 1 = full coverage/several monitoring locations per assessment unit given its size, 0.5 = two or more monitoring locations per assessment unit, and 0 = one monitoring location per assessment unit.

Methodological confidence (ConfM). For coastal fish all assessment units reach level 1 since all monitoring programs included in the evaluation are described in the coastal fish monitoring [guidelines](#).

Table 5. Confidence in the status evaluation of the cyprinids/mesopredators indicator according to the criteria developed within HELCOM for the integrated biodiversity assessment.

Sub-basin	Country	Coastal area name (assessment unit)	Coastal		Time period assessed	Identity of indicator	Monitoring method	Assessment method	ConfA	ConfT	ConfS	ConfM
			area code	Monitoring area/data set								
Bothnian Bay	Finland	Bothnian Bay Finnish Coastal waters	1	NA	NA	NA	NA	NA	NA	NA	NA	NA
Bothnian Bay	Sweden	Bothnian Bay Swedish Coastal waters	2	Kinnbäcksfjärden	2004-2020	Cyprinids	Fisheries independent data	ASCETS	0.5	1	0.5	1
Bothnian Bay	Sweden	Bothnian Bay Swedish Coastal waters	2	Råneå	2002-2020	Cyprinids	Fisheries independent data	ASCETS	0	1	0.5	1
The Quark	Finland	The Quark Finnish Coastal waters	3	NA	NA	NA	NA	NA	NA	NA	NA	NA
The Quark	Sweden	The Quark Swedish Coastal waters	4	Holmön	2002-2020	Cyprinids	Fisheries independent data	ASCETS	0	1	0.5	1
The Quark	Sweden	The Quark Swedish Coastal waters	4	Norrbyn	2002-2020	Cyprinids	Fisheries independent data	ASCETS	1	1	0.5	1
Bothnian Sea	Finland	Bothnian Sea Finnish Coastal waters	5	NA	NA	NA	NA	NA	NA	NA	NA	NA
Bothnian Sea	Sweden	Bothnian Sea Swedish Coastal waters	6	Forsmark	2002-2020	Cyprinids	Fisheries independent data	ASCETS	0.5	1	0.5	1
Bothnian Sea	Sweden	Bothnian Sea Swedish Coastal waters	6	Gavksfjärden	2004-2020	Cyprinids	Fisheries independent data	ASCETS	0.5	1	0.5	1
Bothnian Sea	Sweden	Bothnian Sea Swedish Coastal waters	6	Långvindsfjärden	2002-2020	Cyprinids	Fisheries independent data	ASCETS	0.5	1	0.5	1
Åland Sea	Finland	Åland Sea Finnish Coastal waters	7	NA	NA	NA	NA	NA	NA	NA	NA	NA
Åland Sea	Sweden	Åland Sea Swedish Coastal waters	8	Galtfjärden	2002-2020	Cyprinids	Fisheries independent data	ASCETS	0.5	1	0.5	1
Åland Sea	Sweden	Åland Sea Swedish Coastal waters	8	Lagnö	2002-2020	Cyprinids	Fisheries independent data	ASCETS	1	1	0.5	1
Archipelago Sea	Finland	Archipelago Sea Coastal waters	9	Finbo	2002-2021	Cyprinids	Fisheries independent data	ASCETS	0.5	1	0.5	1
Archipelago Sea	Finland	Archipelago Sea Coastal waters	9	Kumlinge	2002-2021	Cyprinids	Fisheries independent data	ASCETS	1	1	0.5	1
Northern Baltic Sea	Finland	Northern Baltic Proper Finnish Coastal waters	10	NA	NA	NA	NA	NA	NA	NA	NA	NA
Northern Baltic Sea	Sweden	Northern Baltic Proper Swedish Coastal waters	11	Askö	2005-2020	Cyprinids	Fisheries independent data	ASCETS	1	1	0.5	1
Northern Baltic Sea	Sweden	Northern Baltic Proper Swedish Coastal waters	11	Muskö	1992-2020	Mesopredators	Fisheries independent data	ASCETS	1	1	0.5	1
Northern Baltic Sea	Sweden	Northern Baltic Proper Swedish Coastal waters	11	Vaxholm: Askrikefjärden	2016-2020	Cyprinids	Fisheries independent data	Trend	0.5	1	0.5	1
Northern Baltic Sea	Estonia	Northern Baltic Proper Estonian Coastal waters	12	NA	NA	NA	NA	NA	NA	NA	NA	NA
Gulf of Finland	Finland	Gulf of Finland Finnish Coastal waters	13	Brunskär	2002-2020	Cyprinids	Fisheries independent data	ASCETS	1	1	1	1
Gulf of Finland	Finland	Gulf of Finland Finnish Coastal waters	13	Helsinki	2005-2020	Cyprinids	Fisheries independent data	ASCETS	0	0.5	1	1
Gulf of Finland	Finland	Gulf of Finland Finnish Coastal waters	13	Tvärminne	2005-2020	Cyprinids	Fisheries independent data	ASCETS	0.5	1	1	1
Gulf of Finland	Estonia	Gulf of Finland Estonian Coastal waters	14	NA	NA	NA	NA	NA	NA	NA	NA	NA
Gulf of Finland	Russia	Gulf of Finland Russian Coastal waters	15	NA	NA	NA	NA	NA	NA	NA	NA	NA
Gulf of Riga	Estonia	Gulf of Riga Estonian Coastal waters	16	Hiumaa	1991-2020	Cyprinids	Fisheries independent data	ASCETS	1	1	0	1
Gulf of Riga	Latvia	Gulf of Riga Latvian Coastal waters	17	Daugavgriva	2016-2020	Cyprinids	Fisheries independent data	Trend	0	1	0	1
Western Gotland Basin	Sweden	Western Gotland Basin Swedish Coastal waters	18	Kväddfjärden	1998-2020	Mesopredators	Fisheries independent data	ASCETS	0.5	1	0.5	1
Western Gotland Basin	Sweden	Western Gotland Basin Swedish Coastal waters	18	Kväddfjärden	2002-2020	Cyprinids	Fisheries independent data	ASCETS	0.5	1	0.5	1
Western Gotland Basin	Sweden	Western Gotland Basin Swedish Coastal waters	18	Vinö	2007-2020	Cyprinids	Fisheries independent data	ASCETS	0.5	1	0.5	1
Eastern Gotland Basin	Estonia	Eastern Gotland Basin Estonian Coastal waters	19	NA	NA	NA	NA	NA	NA	NA	NA	NA
Eastern Gotland Basin	Latvia	Eastern Gotland Basin Latvian Coastal waters	20	Jurkalne	2016-2020	Cyprinids	Fisheries independent data	Trend	1	1	0	1
Eastern Gotland Basin	Lithuania	Eastern Gotland Basin Lithuanian Coastal waters	21	Curonian lagoon	1998-2020	Cyprinids	Fisheries independent data	ASCETS	1	1	1	1
Eastern Gotland Basin	Lithuania	Eastern Gotland Basin Lithuanian Coastal waters	21	Karkle	2000-2020	Mesopredators	Fisheries independent data	Trend	1	1	1	1
Eastern Gotland Basin	Lithuania	Eastern Gotland Basin Lithuanian Coastal waters	21	Mon/But	1998-2020	Mesopredators	Fisheries independent data	ASCETS	0	1	1	1
Eastern Gotland Basin	Lithuania	Eastern Gotland Basin Lithuanian Coastal waters	21	Smiltyne	2000-2020	Mesopredators	Fisheries independent data	ASCETS	0	1	1	1
Eastern Gotland Basin	Lithuania	Eastern Gotland Basin Lithuanian Coastal waters	21	Šventoji	2000-2020	Mesopredators	Fisheries independent data	ASCETS	1	1	1	1
Eastern Gotland Basin	Sweden	Eastern Gotland Basin Swedish Coastal waters	22	Herrvik	2018-2020	Mesopredators	Fisheries independent data	Trend	0.5	0.5	0	1
Eastern Gotland Basin	Russian	Eastern Gotland Basin Russian Coastal waters	23	NA	NA	NA	NA	NA	NA	NA	NA	NA
Eastern Gotland Basin	Poland	Eastern Gotland Basin Polish Coastal waters	24	NA	NA	NA	NA	NA	NA	NA	NA	NA
Gdansk Basin	Russia	Gdansk Basin Russian Coastal waters	25	NA	NA	NA	NA	NA	NA	NA	NA	NA
Gdansk Basin	Poland	Gdansk Basin Polish Coastal waters	26	Zalew Pucki	2011-2020	Mesopredators	Fisheries independent data	Trend	1	1	1	1
Gdansk Basin	Poland	Gdansk Basin Polish Coastal waters	26	Zalew Wiślany	2011-2020	Cyprinids	Fisheries independent data	Trend	1	1	1	1
Gdansk Basin	Poland	Gdansk Basin Polish Coastal waters	26	Zatoka Pucka Zewnętrzna	2011-2020	Mesopredators	Fisheries independent data	Trend	1	0.5	1	1
Bornholm Basin	Sweden	Bornholm Basin Swedish Coastal waters	27	Hanöbukten	2015-2020	Mesopredators	Fisheries independent data	Trend	0	1	0.5	1
Bornholm Basin	Sweden	Bornholm Basin Swedish Coastal waters	27	Torhamn	2000-2020	Cyprinids	Fisheries independent data	ASCETS	0.5	1	0.5	1
Bornholm Basin	Poland	Bornholm Basin Polish Coastal waters	28	NA	NA	NA	NA	NA	NA	NA	NA	NA
Bornholm Basin	Denmark	Bornholm Basin Danish Coastal waters	29	NA	NA	NA	NA	NA	NA	NA	NA	NA
Bornholm Basin	Germany	Bornholm Basin German Coastal waters	30	NA	NA	NA	NA	NA	NA	NA	NA	NA
Arkona Basin	Sweden	Arkona Basin Swedish Coastal waters	31	Stavstensudde	2018-2020	Mesopredators	Fisheries independent data	Trend	1	0.5	0	1
Arkona Basin	Denmark	Arkona Basin Danish Coastal waters	32	NA	NA	NA	NA	NA	NA	NA	NA	NA
Arkona Basin	Germany	Arkona Basin German Coastal waters	33	NA	NA	NA	NA	NA	NA	NA	NA	NA
Mecklenburg Bight	Germany	Mecklenburg Bight German Coastal waters	34	NA	NA	NA	NA	NA	NA	NA	NA	NA
Mecklenburg Bight	Denmark	Mecklenburg Bight Danish Coastal waters	35	NA	NA	NA	NA	NA	NA	NA	NA	NA
Kiel Bight	Denmark	Kiel Bight Danish Coastal waters	36	NA	NA	NA	NA	NA	NA	NA	NA	NA
Kiel Bight	Germany	Kiel Bight German Coastal waters	37	NA	NA	NA	NA	NA	NA	NA	NA	NA
Belt Sea	Denmark	Belts Danish Coastal waters	38	NA	NA	NA	NA	NA	NA	NA	NA	NA
The Sound	Sweden	The Sound Swedish Coastal waters	39	NA	NA	NA	NA	NA	NA	NA	NA	NA
The Sound	Denmark	The Sound Danish Coastal waters	40	NA	NA	NA	NA	NA	NA	NA	NA	NA
Kattegat	Sweden	Kattegat Swedish Coastal waters, including Limfjorden	41	NA	NA	NA	NA	NA	NA	NA	NA	NA
Kattegat	Denmark	Kattegat Danish Coastal waters, including Limfjorden	42	NA	NA	NA	NA	NA	NA	NA	NA	NA

The confidence in the accuracy of the evaluation (ConfA) is high in 14, medium in 11, and low in 7 of the in total 32 monitoring locations considered. In the locations scoring low for ConfA, there is either short time-series or substantial interannual variation in the indicator value during the assessment period leading to a lower confidence in the evaluation of status. The confidence in the temporal coverage (ConfT) is high in all areas except for the locations of Helsinki (Gulf of Finland, Finland) and Zatoka Pucka Zewnętrzna (Gulf of Gdansk, Poland) due to missing data in one or more of the years in the assessment period. The confidence in spatial representability (ConfS) is only high along the Lithuanian and Polish coasts and low along the southern Swedish coast (Arkona basin) and in Latvian and Estonian coastal waters. In all other areas, ConfS is scored as being intermediate. The methodological confidence (ConfM) is high an all locations evaluated. The integrated confidence considering all four categories varies between high (five assessment units) and intermediate (nine assessment units), but with no clear spatial pattern (Figure 6).

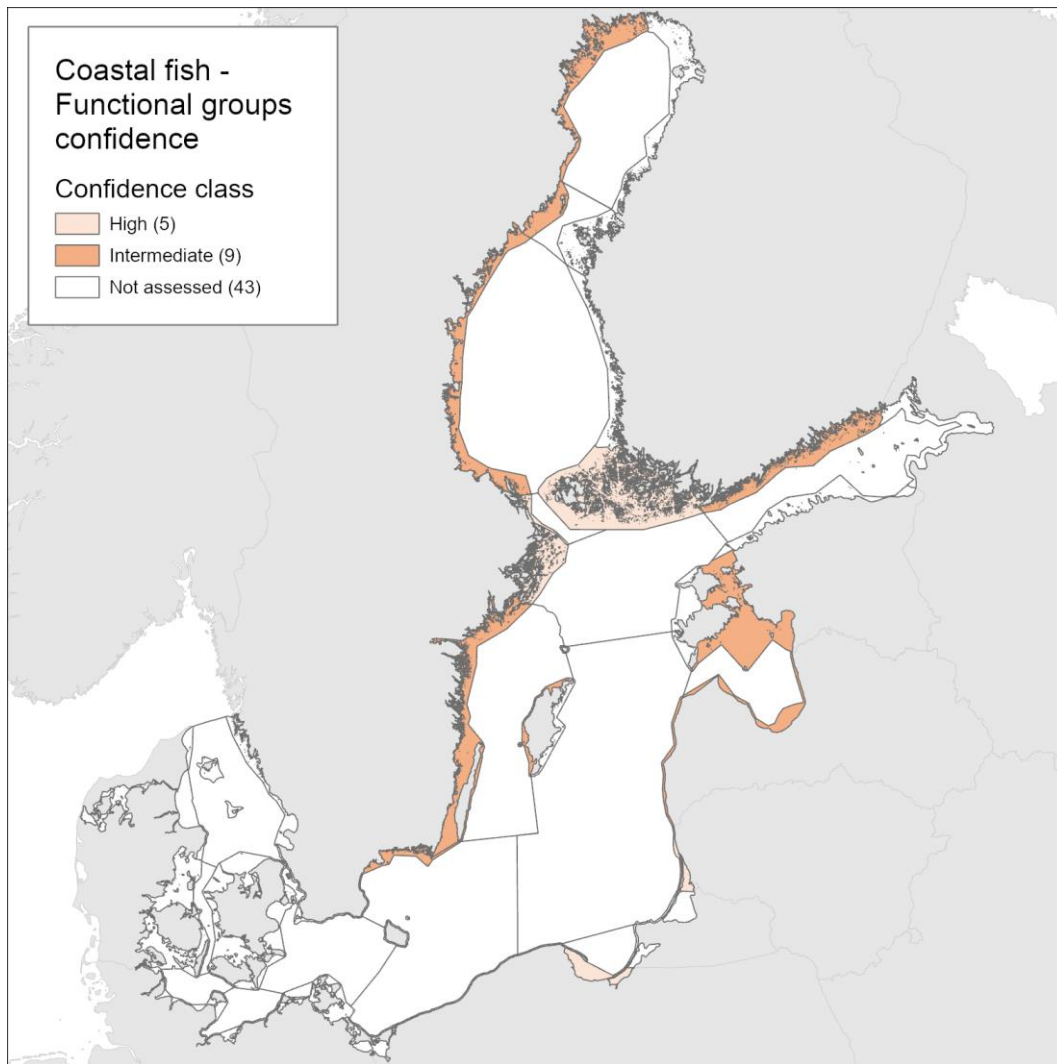


Figure 6. Map of confidence of the current evaluation of the cyprinids/mesopredators indicator. See Table 5 for details.

The confidence concept as developed for the HELCOM integrated biodiversity assessment is not fully applicable to coastal fish as further evaluation of the precision in data and the congruence in status across monitoring locations within assessment units would provide additional information that is needed.

6 Drivers, Activities, and Pressures

The state of key functional groups of coastal fish in the Baltic Sea is influenced by multiple pressures, including climate, eutrophication, exploitation of essential habitats, and in a few areas fishing mortality. Natural processes such as food web interactions and predation from apex predators are also of importance.

The functional groups considered in this indicator are generally heavily affected by the impacts of a changing climate (Olsson *et al.* 2012; Bergström *et al.* 2016b, 2019; Östman *et al.* 2017b) (cf. Chapter 7), including alterations in the food web (Eriksson *et al.* 2009; 2011; Östman *et al.* 2016), the impact of increased water temperature and, for cyprinids in particular, also lowered salinity (Härmä *et al.* 2008; Östman *et al.* 2017b).

Among pressures related to human activities, exploitation of essential habitats (Sundblad *et al.* 2014; Sundblad & Bergström 2014; Kraufvelin *et al.* 2018) impact cyprinids/mesopredators throughout the Baltic, whereas fishing generally affects mainly cyprinids locally in Sweden and Finland (Lappalainen *et al.* 2019; Dahlin *et al.* 2021), and to some extent in the Baltic States and Polish coasts.

The effect of eutrophication on the state of coastal fish communities do mainly affect cyprinids (Härmä *et al.* 2008; Bergström *et al.* 2016b, 2019), and might increase with higher latitude (Östman *et al.* 2017b).

Cyprinids and mesopredatory fish species typically represent lower trophic levels in being planktivores and benthivores. As such, these groups of species are both impacted by bottom-up mechanisms such as eutrophication (Härmä *et al.* 2008; Östman *et al.* 2016) as well as by top-down regulation by piscivorous fish species (Eriksson *et al.* 2011; Baden *et al.* 2012; Casini *et al.* 2012; Östman *et al.* 2016) and apex predators (Östman *et al.* 2012; Hansson *et al.* 2018). Hence, high abundances of cyprinids and mesopredators often characterize ecosystems in an undesirable environmental state.

Natural interactions such as predation pressure from apex predators, foremost cormorants (*Phalacrocorax carbo*), could at least locally impact the state of coastal fish communities (Vetemaa *et al.* 2010; Östman *et al.* 2012; Hansson *et al.* 2018). In some areas the outtake of coastal fish by cormorants exceeds, or is of a similar magnitude, to that of the commercial and recreational fisheries (Östman *et al.* 2013; Hansson *et al.* 2018). The state of groups of mesopredatory fish species such as wrasses, sticklebacks and gobies, and potentially also cyprinids, could be affected by the food web structure in coastal areas and neighbouring ecosystems (Eriksson *et al.* 2011; Baden *et al.* 2012; Casini *et al.* 2012). Especially decreased predation pressure from declining stocks of piscivorous fish species might favour the increase in abundance of mesopredatory fish species (Östman *et al.* 2016). On the other hand, the mesopredators are an important part of the diet of cormorants, which may locally compensate the lack of predatory fish.

Table 6. Brief summary of relevant pressures and activities with relevance to the indicator.

	General	MSFD Annex III, Table 2a
Strong link	Several pressures, both natural and human, acting in concert affect the state of key functional groups of coastal fish. These include climate, eutrophication, fishing, and exploitation and loss of essential habitats. To date, no analyses on the relative importance of these variables have been conducted.	<p><u>Biological</u> <i>Extraction of, or mortality/injury to, wild species</i> (e.g. selective extraction of species, including incidental non-target catches)</p> <p><u>Physical</u> <i>Physical disturbance to seabed</i> (e.g. abrasion and selective extraction)</p> <p><i>Physical loss</i> (e.g. sealing) <i>Changes to hydrological processes</i> (e.g. significant changes in thermal and/or salinity regime)</p> <p><u>Substances</u> <i>Inputs of nutrients</i> (e.g. inputs of fertilisers and other nitrogen and phosphorus-rich substances)</p>
Weak link	There might also be effects of hazardous substances and non-indigenous species on the state of coastal fish key functional groups	<p><u>Substances</u> <i>Input of other substances</i> (e.g. synthetic substances, non-synthetic substances, radionuclides)</p> <p><u>Biological</u> <i>Input or spread of non-indigenous species</i></p>

7 Climate change and other factors

The functional groups considered in this indicator are generally heavily affected by the impacts of a changing climate (Olsson *et al.* 2012; Bergström *et al.* 2016b, 2019; Östman *et al.* 2017b), including alterations in the food web (Eriksson *et al.* 2009; 2011; Östman *et al.* 2016), the impact of increased water temperature and, for cyprinids in particular, also lowered salinity (Härmä *et al.* 2008; Östman *et al.* 2017b) (cf. section 6 of this report).

8 Conclusions

8.1 Future work or improvements needed

Due to the presence of natural environmental gradients across the Baltic Sea, and the rather local appearance of coastal fish communities (and hence their differing structures and responses to environmental change), the spatial coverage of monitoring should be improved in some areas in order to enhance the confidence of the evaluation outcome. When designating new potential monitoring programmes, it should be considered that the levels of direct human impact on the coastal fish communities in many of the existing monitoring areas are low, and future locations should also include more heavily affected areas.

In addition, as a multitude of factors with natural environmental gradients in the Baltic Sea potentially impact coastal fish communities and species, the magnitude of importance of different factors in different coastal areas should be understood. A more mechanistic understanding of how pressures impact upon coastal fish in local contexts will enable managers to take relevant measures to halt declining trends of coastal fish species in some coastal areas.. More specifically, the role of fishing (both commercial and recreational) and natural predation needs further investigation.

9 Methodology

This indicator uses two different approaches for evaluating whether Good Status is achieved. The approach used depends on the availability of data. If certain criteria are met, the ASCETS method is used (Östman *et al.* 2020). If not, the trend-based approach is used.

The methodology and basis of the indicator evaluation is provided below.

9.1 Scale of assessment

Due to the local appearance of typical coastal fish populations, status evaluations of coastal fish communities are representative of rather small geographical scales, however, there is scope to further interrogate the citizen science monitoring data to try and develop a mesopredator index independent of the abundances of flounder and eelpout, which are currently used in the "Abundance of coastal fish species" indicator. In this evaluation the HELCOM assessment unit scale 3 'Open sub-basin and coastal waters' has been applied. The indicator is not evaluated for the open sea sub-basins since the species in focus are coastal.

Evaluations for both indicators were carried out for 16 coastal HELCOM assessment units, but in two Swedish units the time-series was too short to allow for an evaluation against a quantitative threshold value. The number of units evaluated is currently restricted by the availability of monitoring data.

For the integration of status across species and monitoring locations within assessment units, the One-Out-All-Out principle is applied (Dierschke *et al.* 2021).

The assessment units are defined in the [HELCOM Monitoring and Assessment Strategy Annex 4](#).

9.2 Methodology applied

ASCETS approach

Coastal fish datasets must meet certain criteria in order to be able to apply an evaluation of good status using the ASCETS approach:

1. The time period used to determine the reference period should cover a minimum number of years that is twice the generation time of the species most influential in the indicator evaluation. This is to ensure that the influences of strong year classes are taken into account. For coastal fish, this is typically about ten years. In this evaluation, the time period used to determine the reference period against which good status is evaluated spans the years 1998 to 2015, with varying numbers of years depending on data availability for each time series.
2. Before evaluating good status, it should be decided whether or not the reference period reflects good status. If a previous status evaluation exists from HOLAS 2, the reference period is assigned the same status as the assessment period in HOLAS 2 (2011-2016). If a previous status evaluation does not exist, this can be

done by using data dating back earlier than the start of the period used to determine the reference period, using additional information, or by expert judgment. For example, if data from time periods preceding the period used for determining the reference period have much higher indicator values, the reference might represent not good status (in case of an indicator where higher values are indicative of a good environmental state) or good status (in case of an indicator where higher values are indicative of an undesirable state).

The ASCETS method (Östman *et al.* 2020) offers a refined approach to infer structural changes in indicator values over time and establish threshold values for the state during a reference period based on the observed variation in indicator values. ASCETS also gives estimates on the confidence of an apparent change in state of indicator values between a reference period and an assessment period. Thus, by applying ASCETS to time series data, it is possible to derive threshold values for addressing structural changes in indicator values over time and a developed evaluation of the confidence of the derived current indicator state relative to previous indicator values. To determine the status of the indicator, the ASCETS method first derives a bootstrapped distribution of median values from a time series of observed indicator values during a reference period. Specific threshold values for changes in indicator state is set based on the Xth and XXth percentile values of the bootstrapped distribution. For functional groups, the percentiles are 5 and 95/98 percent (depending on the status of the reference period, see below), representing the confidence interval of median indicator values. In this way, the derived boundaries of the confidence interval can function as threshold values for a change in state per assessment unit of each species. Because ASCETS bootstraps median indicator values during the reference period it is possible that one or several observed indicator values during the reference period will fall outside of the 95% confidence interval, because the bootstrapping reduces the influence of what may be large sampling errors. Second, the bootstrapped median indicator value during the assessment period is evaluated in relation to the threshold values derived from the reference period depending on how much of the bootstrapped median distribution from the assessment period that falls below, within, or above the Xth and XXth percentiles (cf. Figure 3 and decision tree in Figure 7):

1. In situations where the baseline state reflects good status, the median of the years in the assessment period should be above the 5th percentile and below the 95th percentile to reflect good status.
2. In situations where the baseline state reflects not good status, in order to reflect good status, the median of the years in the assessment period should be above the 98th percentile if the baseline status is indicative of too low abundances, and below the 5th percentile if the baseline status is indicative of too high abundances.

Trend-based approach

If the requirements for defining quantitative baseline conditions are not met (e.g. short time series), then a trend-based evaluation should be used. All available data starting from year 2014 is included in trend analyses.

In the trend based approach, good status is defined based on the direction of the trend compared to the desired direction of the indicator over time (cf. Figure 4). Where the first years in the evaluated time series represent good status, the trend of the indicator over time should not exhibit any direction in order to reflect good status. If, on the other hand, the first years of the evaluated time series represent not good status, the trend should be in the desired direction to reflect good status. The significance level for these trends should be $p < 0.1$.

Decision tree for evaluation using coastal fish community structure

The assessment protocol is found in figure 7.



Figure 7. Decision tree for status evaluation using coastal fish community structure. ASCETS approach (top figure) and the trendbased approach (bottom figure) are presented.

Data analyses

The data used for the evaluations are derived from fishery independent monitoring. The analyses are based on catch per unit effort (CPUE) data from annual averages of all sampling stations in each area. Individuals smaller than 12 cm (Nordic Coastal multimesh nets) or 14 cm (other net types) were excluded from the evaluation in order to only include species and size-groups suited for quantitative sampling by the method. Abundance is calculated as the number of individuals of the species included in the indicator per unit effort (CPUE).

9.3 Monitoring and reporting requirements

Monitoring methodology

The HELCOM common monitoring on coastal fish is described on a general level in the HELCOM Monitoring Manual in the [sub-programme: Coastal fish](#).

The HELCOM common monitoring on coastal fish is described in [guidelines](#) that were adopted in 2014 and updated in 2019.

Current monitoring

The monitoring activities relevant to the indicator that are currently carried out by HELCOM Contracting Parties are described in the HELCOM Monitoring Manual in the Monitoring Concepts table as well as in the [guidelines for coastal fish monitoring](#).

Sub-programme: Coastal fish

[Monitoring Concepts table](#)

Coastal fish monitoring is rather widespread in the Baltic Sea, and at present covers 32 of the in total 42 'scale 3 HELCOM assessment units'. Coastal areas that lack coastal fish monitoring includes Russia and Germany (in total 7 assessment units) where there is no current and official monitoring program for coastal fish, two assessment units in Finland (Åland Sea Finnish coastal waters and Northern Baltic Proper Finnish coastal waters) and one in Denmark (Kiel Bight Danish coastal waters). The current monitoring where information cyprinids/mesopredators can be extracted to date is less extensive, covering 14 assessment units.

The current monitoring of coastal fish in the Baltic Sea represents a minimum level of effort and serves as a first step for evaluating the status of coastal fish communities.

The current monitoring likely yields insights into major and large-scale changes in coastal fish communities in the Baltic Sea, but unique and departing responses are possible in some areas.

In Estonia and Latvia, coastal fish monitoring is carried out at several locations, but the evaluation has only been made for one location in Estonia and two in Latvia. In Denmark, no data is available to support the cyprinids/mesopredators, and the Finnish commercial

catch data is not applicable for assessing status of non-targeted fish species. In Germany, there is no coordinated monitoring program for coastal fish, but a [project](#) aiming to establish such a program was initiated in 2020 in the coastal areas of Schleswig-Holstein.

10 Data

The data and resulting data products (e.g. tables, figures and maps) available on the indicator web page can be used freely given that it is used appropriately and the source is cited.

[Result: Abundance of coastal fish key functional groups – integrated result](#)

[Result: Abundance of coastal fish key functional groups – cyprinids](#)

[Result: Abundance of coastal fish key functional groups – mesopredators](#)

[Data: Abundance of coastal fish key functional groups](#)

Data are typically collected annually in August by national and regional monitoring programmes (cf. HELCOM (2019) for details). For future updates of this evaluation, data should be collected in each location on an annual basis.

A few time series of coastal fish began in the 1970s (Olsson *et al.* 2012), whereas others were started in the 1980s and the mid-1990s (HELCOM 2019). In Finland and Sweden, a new coastal fish monitoring programme with a higher spatial resolution was established in the early 2000s, and in Poland monitoring data is typically available from the mid 2010s. For more information, see HELCOM 2019.

The raw data on which this evaluation is based, are stored in national databases. Each country has its own routines for quality assurance of the stored data. From 2017, each country calculates indicator values for their monitoring locations from the raw data from fish monitoring. The indicator data and values are then during the first half of the year uploaded to the HELCOM database for coastal fish core indicators, COOL as hosted by the HELCOM secretariat. Indicator data for status evaluations are extracted from the COOL database, and the evaluation undertaken by the lead country (Sweden) according to the assessment protocol outlined in this report.

Data source

Coastal fish monitoring is coordinated within the HELCOM [FISH-PRO III](#) expert network. The network compiles data from various sources of data for coastal fish in Finland, Estonia, Latvia, Lithuania, Poland, Germany, Denmark and Sweden (HELCOM 2019). In Germany, there is no coordinated monitoring program for coastal fish, but a [project](#) aiming to establish such a program was initiated in 2020 in the coastal areas of Schleswig-Holstein.

The institutes responsible for sampling are: Natural Resources Institute Finland (Luke) (Finland), Provincial Government of Åland Islands (Finland), Estonian Marine Institute (Estonia), University of Tartu (Estonia), Institute of Food Safety, Animal Health and Environment "BIOR" (Latvia), Nature Research Center (Lithuania), Klaipeda University

(Lithuania), National Marine Fisheries Research Institute, Gdynia (Poland), Association Fish and Environment Mecklenburg-Vorpommern e.V. (Germany), University of Rostock (Germany), National Institute of Aquatic Resources, Technical University of Denmark (Denmark), Department of Aquatic Resources, Swedish University of Agricultural Sciences (Sweden).

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12 Archive

This version of the HELCOM core indicator report was published in April 2023:

The current version of this indicator (including as a PDF) can be found on the [HELCOM indicator web page](#).

Earlier versions of the core indicator report include:

[Abundance of coastal fish key functional groups HELCOM core indicator 2018](#) (pdf)

[HOLAS 2 component - Core indicator report – web-based version July 2017](#) (pdf)

[Core indicator report – web-based version October 2015](#) (pdf)

[Extended core indicator report – outcome of CORESET II project \(pdf\) \(2015\)2013 Indicator report \(pdf\)](#)

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